

PDF hosted at the Radboud Repository of the Radboud University Nijmegen

The following full text is a publisher's version.

For additional information about this publication click this link.

<http://hdl.handle.net/2066/202539>

Please be advised that this information was generated on 2021-06-24 and may be subject to change.



Nutrient enrichment changes the nature of invertebrate food webs in raised bog pools

G.A. van Duinen^{1,2,3}, K. Vermonden², A.M.T. Brock¹, R.S.E.W. Leuven², A.J.P. Smolders⁴, G. van der Velde³, W.C.E.P. Verberk^{1,3} and H. Esselink^{1,3}

¹ Bargerveen Foundation, Radboud University Nijmegen, P.O. Box 9010, 6500 GL Nijmegen, The Netherlands
Phone: +31 24 3653296, Fax: +31 24 3653030, e-mail: G.vanDuinen@science.ru.nl

² Department of Environmental Science, Radboud University Nijmegen, P.O. Box 9010, 6500 GL Nijmegen, The Netherlands

³ Department of Animal Ecology and Ecophysiology, Radboud University Nijmegen, P.O. Box 9010, 6500 GL Nijmegen, The Netherlands

⁴ Department of Aquatic Ecology and Environmental Biology, Radboud University Nijmegen, P.O. Box 9010, 6500 GL Nijmegen, The Netherlands

Summary

Drainage and increased atmospheric nitrogen and sulphur deposition results in increased nutrient availability in naturally nutrient-poor bogs. To study if increased nutrient availability changes the bog food web, stable isotope ratios of carbon and nitrogen of basal food sources (living plants and dead organic matter) and invertebrates were compared between bog pools, differing in nutrient availability, in Estonia and the Netherlands. For part of the invertebrate community, trophic relations could not be inferred from measured basal food sources. One or more basal carbon sources with a low stable isotope ratio for carbon still need to be identified. This unidentified carbon source becomes less important with increasing nutrient availability.

Key index words: eutrophication, nutrients, fauna, invertebrates, food web

Introduction

Intact raised bogs are extremely nutrient-poor and bog plants have correspondingly low nutritional value. Due to atmospheric nitrogen and sulphur deposition nutrient availability has increased in Western European raised bog pools and a simultaneous change in the invertebrate species composition has been documented (Desrocher and Van Duinen, 2006; Van Duinen *et al.*, 2006). SO_4 can lead to an increased release of NH_4 and PO_4 from the peat substrate in bog pools (Lamers *et al.*, 1998). Increased nutrient availability in raised bogs leads to a higher cover of vascular plants such as *Molinia caerulea* and *Betula* spp. and an elevated N-content of *Sphagnum* mosses (Lamers *et al.*, 2000). As a result, both the input and nutritional value of dead organic material in bog pools increase. Additionally, algal growth rate and nutrient content are increased. These changes in quantity and quality of food sources at the start of the aquatic food web may underlie the observed changes in invertebrate species composition. This paper addresses the question whether the increased nutrient availability has caused changes in the importance of basal food sources that may have led to cascading effects through the invertebrate food web in bog pools.

Material and methods

Stable isotope ratios of carbon and nitrogen of basal food sources and aquatic invertebrates were compared between nine bog pools differing in nutrient availability. Three bog pools were sampled in Nigula bog, Southwest Estonia,

between 9 and 23 September 2002. In the Netherlands, six peat cuttings and inundated peat extraction fields were sampled in five raised bog remnants in total between 7 October and 15 November 2002. At each site surface water and sediment pore water samples were collected for analysis of nutrient availability (Table 1). For further details about water quality analyses see Van Duinen *et al.* (2006).

At all sampling sites plants (filamentous algae, mosses and vascular plants) and aquatic macroinvertebrates were collected. Settled dead organic matter was collected from the peat bottom by means of a plankton net with mesh size 45 μm . Zooplankton was collected from the open water by means of this plankton net and light traps. Gut contents were not removed from invertebrates. The collected material was sorted, washed with demineralised water and kept in a fridge until identification. Identified material was dried for 24 hours at 70°C and subsequently ground, using liquid nitrogen. Carbon and nitrogen isotopic composition was determined. Average reproducibilities of duplo and triplo measurements were <0.2‰. Stable isotope data are presented as the relative difference between the ratios of the sample and the standards, using the following formula:

$$\delta R = [(R_{\text{sample}} / R_{\text{standard}}) - 1] \times 1000$$

where $R = {}^{13}\text{C}/{}^{12}\text{C}$ or ${}^{15}\text{N}/{}^{14}\text{N}$. $\delta^{13}\text{C}$ or $\delta^{15}\text{N}$ is the per mille (‰) deviation of the sample from their isotope standards, that are Pee Dee belemnite for $\delta^{13}\text{C}$ and atmospheric N_2 for $\delta^{15}\text{N}$.



Table 1. Average (\pm sd) pH and concentrations of o-PO₄ and NO₃+NH₄ (μ mol/l) of surface water and sediment pore water at the nine sampling sites.

Site	Surface water			Sediment pore water		
	pH	o-PO ₄	NO ₃ +NH ₄	pH	o-PO ₄	NO ₃ +NH ₄
<i>Estonia</i>						
Nigula 1	4.2 \pm 0.3	0.2 \pm 0.1	13 \pm 14	4.9 \pm 0.2	0.2 \pm 0.1	5.4 \pm 4.0
Nigula 2	4.0 \pm 0.0	0.1 \pm 0.1	12 \pm 5.6	4.8 \pm 0.7	0.2 \pm 0.1	15 \pm 16
Nigula 3	3.9 \pm 0.1	0.3 \pm 0.0	11 \pm 9.0	4.5 \pm 0.2	0.2 \pm 0.1	21 \pm 14
<i>The Netherlands</i>						
Haaksbergerveen 1	4.2 \pm 0.2	0.8 \pm 0.4	14 \pm 12	4.8 \pm 0.5	1.3 \pm 1.2	46 \pm 21
Haaksbergerveen 2	4.2 \pm 0.1	2.2 \pm 2.8	13 \pm 7.4	5.5 \pm 0.5	1.7 \pm 0.5	40 \pm 14
Korenburgerveen	3.7 \pm 0.1	0.9 \pm 0.2	11 \pm 4.1	5.1 \pm 0.9	0.8 \pm 0.1	21 \pm 11
Reigerplas	4.2 \pm 0.5	0.4 \pm 0.3	8.3 \pm 2.8	5.0 \pm 0.0	3.2 \pm 2.3	110 \pm 131
Mariapeel	4.0 \pm 0.1	3.4 \pm 17	23 \pm 21	5.0 \pm 0.4	18 \pm 19	201 \pm 117
Bargerveen	3.8 \pm 0.1	5.6 \pm 8.2	20 \pm 16	4.5 \pm 0.5	82 \pm 102	213 \pm 219

Results and Discussion

Different invertebrate species collected within the same bog pool differ in their $\delta^{13}C$, indicating differential use of carbon sources, and in their $\delta^{15}N$, indicating the presence of various trophic levels in the invertebrate community (Figure 1). For dead organic matter and plants, as well as most invertebrate species trophic levels in the food web inferred from their $\delta^{15}N$ corresponded with common ecological knowledge.

In food web studies generally an enrichment of 0 to 1‰ (less negative) for $\delta^{13}C$ between trophic levels is found (McCutchan *et al.*, 2003). Only for the two most nutrient rich sites (Bargerveen and Mariapeel) all collected invertebrate species appeared to be similar or enriched in $\delta^{13}C$ compared to basal food sources (dead organic matter and living plants). In the three Estonian sites and the less nutrient-rich Dutch sites many invertebrates were depleted (more negative) in $\delta^{13}C$ compared to the analysed basal

food sources. For example, in pool Nigula 1, $\delta^{13}C$ of dead organic matter and living plants ranged from -28.3 to -24.3‰, whereas $\delta^{13}C$ of invertebrates ranged from -34.8 to -24.2‰ (Fig. 1). Although consumers can be depleted in $\delta^{13}C$ relative to their food sources by more than 2‰ (McCutchan *et al.*, 2003), the depletion was not consistent for all sampling sites. The difference between the average $\delta^{13}C$ of all basal food sources and the average $\delta^{13}C$ of all invertebrates (=average trophic shift) correlated strongly and positively to nutrient concentrations of surface water ($p < 0.025$) and sediment pore water ($p < 0.001$; Fig. 2).

For the nutrient poor bog pools still one or more important basal carbon sources with $\delta^{13}C < -30$ ‰ need to be identified. The still unidentified basal carbon sources become less important with increasing nutrient availability. In a study on lake zooplankton, Grey *et al.* (2000) concluded that the relative importance of allochthonous sources of organic carbon decreases with increasing lake

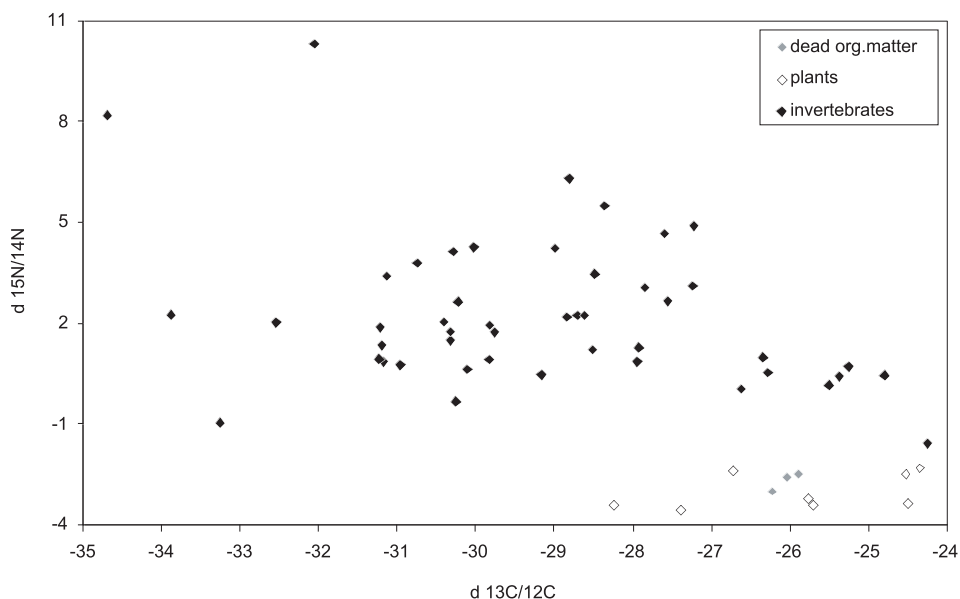


Figure 1. Scatter plot of $\delta^{15}N$ against $\delta^{13}C$ of dead organic matter, plants (vascular plants, mosses and filamentous algae) and aquatic invertebrates of the intact raised bog pool Nigula 1.

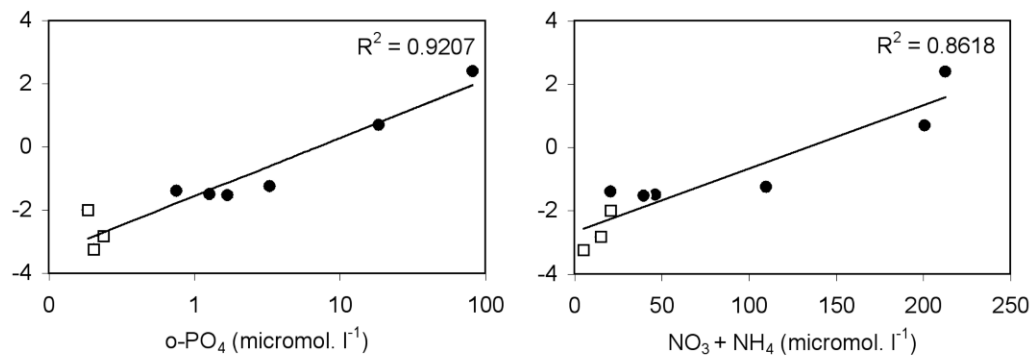


Figure 2. Relationship between the o-PO_4 and $\text{NO}_3 + \text{NH}_4$ concentrations in sediment pore water and the average trophic shift for the nine sampling sites. Open squares for Estonian sites and filled circles for Dutch sites.

trophy. In the case of the pools sampled in Nigula bog, these allochthonous food sources could be pollen, seeds or leaves of *Betula* and *Pinus* trees, that are scattered in the surrounding bog, and terrestrial insects, for example ants and butterflies, falling and drowning in the pool. The $\delta^{13}\text{C}$ of *Betula* leaves and terrestrial invertebrates collected here was $> -29,6\text{‰}$. Alternatively, in oligotrophic lakes in which phytoplankton production is limited, zooplankton diets are supported by planktonic heterotrophs and detritus via the microbial pathway (Hessen *et al.*, 1990; Jones, 1992). Methanotrophic or chemoautotrophic bacteria are known to have a $\delta^{13}\text{C} < -40\text{‰}$ and are assimilated by invertebrates such as chironomid larvae (Kiyashko, 2004; Grey and Deines, 2006) and zooplankton (Taipale *et al.*, 2007).

We suggest that only in the most nutrient enriched raised bog pools the quality of dead organic matter and/or the abundance and quality of algae are sufficient to support the whole invertebrate community. In the pristine or less nutrient enriched pools one or more basal carbon sources with $\delta^{13}\text{C} < -30\text{‰}$ are more important. Thus, nutrient enrichment in bog pools changes the importance of basal carbon sources in the invertebrate food webs.

Acknowledgements

Nigula Nature Reserve Administration, Natuurmonumenten and Staatsbosbeheer are acknowledged for giving permission to collect samples in their reserves. J. Javoiš, J. Kuper, T. Peeters, M.J. Sanabria, M. Smits, M. van der Weijden and Y. Zhuge assisted with field work and identification of collected fauna. J. Eygensteyn performed all isotope measurements. This research project is part of the national research programme 'Survival Plan for Woodland and Nature', funded by the Dutch Ministry of Agriculture, Nature and Food Quality.

References

- Desrochers, A. and Van Duinen, G.A. (2006). Peatland Fauna. In R.K. Wieder and D.H. Vitt (eds.), *Boreal Peatland Ecosystems. Ecological Studies* **18**, 67-100. Springer-Verlag, New York.
- Grey, J., Jones, R.I. and Sleep, D. (2000). Stable isotope analysis of the origins of zooplankton carbon in lakes of differing trophic state. *Oecologia* **123**, 232-240.
- Grey, J. and Deines, P. (2006). Differential assimilation of methanotrophic and chemoautotrophic bacteria by lake chironomid larvae. *Aquat. Microb. Ecol.* **40**, 61-66.
- Hessen, D.O., Andersen, T. and Lyche, A. (1990). Carbon metabolism in a humic lake: Pool sizes and cycling through zooplankton. *Limnology and Oceanography* **35**, 84-99.
- Jones, R.I. (1992). The influence of humic substances on lacustrine planktonic food chains. *Hydrobiologia* **229**, 73-91.
- Kiyashko, S.I., Imbs, A.B., Narita, T., Svetashev, V.I. and Wada, E. (2004). Fatty acid composition of aquatic insect larvae *Stictochironomus pictulus* (Diptera: Chironomidae): evidence of feeding upon methanotrophic bacteria. *Comparative Biochemistry and Physiology* **B 139**, 705-711.
- Lamers, L.P.M., Tomassen, H.B.M. and Roelofs, J.G.M. (1998). Sulfate induced eutrophication and phytotoxicity in freshwater wetlands. *Environmental Science and Technology* **32**, 199-205.
- Lamers, L.P.M., Bobbink, R. and Roelofs, J.G.M. (2000). Natural nitrogen filter fails in polluted raised bogs. *Global Change Biology* **6**, 583-586.
- McCutchan Jr, J.H., Lewis Jr, W.M., Kendall, C. and McGrath, C.C. (2003). Variation in trophic shift for stable isotope ratios of carbon, nitrogen, and sulfur. *Oikos* **102**, 378-390.
- Taipale, S., Kankaala, P. and Jones, R.I. (2007). Contributions of different organic carbon sources to *Daphnia* in the pelagic foodweb of a small polyhumic lake: Results from mesocosm DI^{13}C -additions. *Ecosystems* **10**, 757-772.
- Van Duinen, G.A., Timm, T., Smolders, A.J.P., Brock, A.M.T., Verberk, W.C.E.P. and Esselink, H. (2006). Differential response of aquatic oligochaete species to increased nutrient availability – a comparative study between Estonian and Dutch raised bogs. *Hydrobiologia* **564**, 143-155.