Risk analysis of non-native Curly Waterweed (Lagarosiphon major) in the Netherlands

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Radboud University Nijmegen, Institute for Water and Wetland Research Department of Environmental Sciences, FLORON & Roelf Pot Research and Consultancy

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Title: Risk analysis of non-native Curly Waterweed (*Lagarosiphon major*) in the Netherlands


Cover photo: Dense vegetation of Curly Waterweed (*Lagarosiphon major*) in a ditch near Ter Apel, the Netherlands (Photo: R. Pot)

Project manager: Dr. R.S.E.W. Leuven, Department of Environmental Science, Institute for Water and Wetland Research, Radboud University Nijmegen, Heyendaalseweg 135, 6525 AJ Nijmegen, the Netherlands, e-mail: r.leuven@science.ru.nl

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Summary

Curly Waterweed (*Lagarosiphon major*) is an aquatic plant, non-native to the Netherlands. After first being observed in 2003 in the Soest administrative area, *L. major* has been recorded in the southern and northern provinces. Previously, there was a lack of knowledge regarding the probability of arrival, establishment and spread, (potential) impacts and options for management of *L. major* and management options in the Netherlands. This report is the synthesis of results obtained from a literature study, field observations and expert consultation that address this knowledge gap in the form of a knowledge document. The knowledge document was used to assess the ecological risk using the Belgian Invasive Species Environmental Impact Assessment (ISEIA) protocol. Socio-economic and public health risks were assessed separately as these risk categories do not form part of the ISEIA protocol. Recommendations were then made regarding management options relevant to the situation found in the Netherlands.

The probability of *L. major* arriving in the Netherlands is determined largely by the plant trade. Circa 20,000 units of *L. major* were imported to the Netherlands in 2006. A number of internet websites were found that featured traders who were advertising *L. major* for sale within the Netherlands under the Dutch common name ‘Gekroesde waterpest’. Once bought there is a risk that hobbyist will dispose of excess plants to freshwater bodies. We predict that without management intervention, *L. major* introductions will continue, leading to potential increases in its distribution within the Netherlands. After considering the above information the probability of arrival was judged to be high.

*L. major* was first recorded in ditches in Soest when it was seen to have established in 4 kilometre squares. However, the plants were located at a single location around the junction of these squares. The rate of establishment peaked in the years 2007 and 2008. Since 2008 the number of reported records has decreased compared with preceding years. The current recorded distribution of *L. major* in the Netherlands is mainly characterised by isolated populations of plants. Nine years after being first recorded, the probability of establishment within the Netherlands was judged to be low due to 1) the limited recorded distribution of *L. major*, 2) possible lack of vectors at locations where the species was introduced and 3) minimum temperature tolerance (> 10 °C) in relation to current water temperatures (< 10°C) during winter periods (species-environment mismatch).

Due to the lack of female plants in the Netherlands, reproduction and spread are facilitated by the detachment of small fragments and lateral branches. These fragments subsequently become rooted, developing into new plants. Reproduction through the spread of fragments facilitated by vectors is, therefore, of primary importance in the Netherlands. Vectors can be ordered in terms of importance: the plant trade, hobbyists, boats and water flow (high); weed harvesters (medium-high), fishing equipment (medium); vehicles, large aquatic birds (low). The probability of spread within the Netherlands was judged to be high.
Four factors are considered as part of the (ISEIA) protocol: dispersion potential and invasiveness, colonisation of habitats with high conservation values, adverse impacts on native species and alteration of ecosystem functions.

- Dispersion potential and invasiveness: *L. major* has a strong reproductive potential, can disperse via hydrochory and the level of imports for use in the plant trade remain high. The dispersal of *L. major* away from its initial points of introduction in the Netherlands has been limited and its distribution is characterised by isolated populations. This may be due to isolation or low water flow of the colonised water bodies and a lack of secondary dispersal vectors.

- Colonisation of high conservation value habitats: *L. major* is found in and around urban areas and no records exist in high conservation value habitats (i.e. habitat types in accordance with EU Habitats Directive or Bird Directive) in the Netherlands. However, there is potential that a protected nature area and peat-land with similar characteristics to the H3150 EU Habitats Directive type around Soest may be colonised in the future.

- Adverse impacts to native species: While there are many examples of impacts of *L. major* on native species observed in other countries, until now no effects have been observed in the Netherlands. However, future changes to habitat resulting from e.g. climate change, may increase (potential) risks in the future.

- Alteration to ecosystem functions: There is limited evidence demonstrating negative impacts on the functioning of ecosystems in the Netherlands. However, in other countries, featuring milder climates, where *L. major* has become more widely established, negative impacts on ecosystem functioning have been extensive. Future changes to habitat resulting from e.g. climate change (e.g. increase in minimum water temperature during winter periods), may result in a revision of this risk score in the future.

Using the ISEIA protocol for a risk assessment within the context of the Netherlands, *L. major* was rated as a medium risk species for ecological impacts.

*L. major* is classified in the medium risk category using the ISEIA protocol, however, there is a large body of evidence from abroad that demonstrates the high level of impact that can occur on native species and ecosystem functions if *L. major* becomes more wide spread. Currently the recorded distribution of *L. major* in the Netherlands remains limited. Possible reasons for this could be the minimum temperature tolerance for survival of *L. major* (10 °C) and the lack of dispersal vectors at locations where *L. major* has been recorded. Countries where *L. major* has had a high level of impact on native species and ecosystem functions feature milder climates than the Netherlands. Future habitat changes due to climate change may result in a wider distribution of *L. major* (depending on potential management interventions e.g. isolation of the species or weed cutting) and re-classification of the species to a high risk category.

Due to limited distribution of *L. major* in the Netherlands, the current socio-economic impact of the species is low. However, potential future changes as a result of a rise in
water temperature due to climate change, may increase the suitability and area of *L. major* habitat leading to increased socio-economic impact.

There was no information found concerning the public health effects of *L. major* during the literature study or in communications with project partners.

Banning of sale of *L. major* via the plant trade and creating consumer awareness are the best options for preventing new introductions and controlling further spread. Based on current dispersion and potential invasiveness and risk it is recommended that *L. major* plants are sold with information that informs buyers of their potential invasive nature and the circumstances within which the plant can safely be used. Once established, the management of plants is challenging. Managers may wish to consider observing the dispersal potential of individual populations of *L. major* prior to instigating active measures. If populations become problematic, isolation of plants may be considered as this will facilitate the elimination of the species. Costs and the risk of facilitating dispersal through fragmentation, together with the limited dispersal of *L. major* observed in the Netherlands to date, count against the early implementation of weed cutting as a control measure.
1. Introduction

1.1 Background and problem statement

Curly Waterweed (*Lagarosiphon major*) is a member of a genus of plants that are all endemic to different parts of Southern Africa. *L. major* is a well-defined species and is the only species of the genus *Lagarosiphon* that has been cultivated and introduced elsewhere (Symoens & Triest, 1983). *L. major* was first recorded in the Netherlands in the municipality of Soest around 2003 (Van Valkenburg & Pot, 2008). Over the past decade, this plant species was also recorded at locations in the southern and northern provinces. Previously, there was a lack of knowledge regarding the probability of arrival, establishment and spread, (potential) impacts and options for management of *L. major* in the Netherlands.

To support decision making with regard to the design of measures to prevent ecological, socio-economical and public health effects, the Invasive Alien Species Team of the Netherlands Food and Consumer Product Safety Authority (Ministry of Economic Affairs, Agriculture and Innovation) has asked to carry out a risk analysis of *L. major*. The present report assesses relevant available knowledge and data which is subsequently used to perform a risk analysis of this species.

1.2 Research goals

The major goals of this study are:

- To perform a risk analysis based on the probability of arrival, establishment and spread, endangered areas, the (potential) ecological, socio-economic and public health impacts of *L. major* in the Netherlands.

- To assess the dispersion, invasiveness and (potential) ecological effects of *L. major* in the Netherlands using the Belgian Invasive Species Environmental Impact Assessment (ISEIA) protocol.

- To describe effective risk management options for control of spread, establishment and negative effects of *L. major* into and within the Netherlands.

1.3 Outline and coherence of research

The problem statement and research goals in order undertake a risk analysis of *L. major* in the Netherlands have been described above. Chapter 2 gives the methodological framework of the project, describes the Belgian ISEIA protocol and approaches to assess socio-economic risks and public health risks, and analyses management approaches applicable in the Netherlands. Chapter 3 describes the results of the risk assessment, assesses the probability of arrival, establishment and spread, summarizes the results of the literature study of socio-economic and public health risks and analyses risk management options. Chapter 4 discusses gaps in knowledge and uncertainties, other available risk analyses and explains differences between risk classifications.
Chapter 5 draws conclusions and gives recommendations for further research. An appendix containing background information in the form of a knowledge document completes this report. The coherence between various research activities and outcomes of the study are visualised in a flow chart (Figure 1.1).

Figure 1.1: Flowchart visualising the coherence of various components of the risk analysis of Curly Waterweed (*Lagarosiphon major*) in the Netherlands. Chapter numbers are indicated in brackets.
2. Methods

2.1 Components of the risk analysis

The risk analysis of Curly Waterweed (*Lagarosiphon major*) in the Netherlands was comprised of analyses of probability of arrival into and within the Netherlands, establishment and spread within the Netherlands, endangered areas and an ecological risk assessment using the Belgian Invasive Species Environmental Impact Assessment (ISEIA), developed by the Belgian Biodiversity Platform (Branquart, 2007; ISEIA, 2009). Separate assessments of socio-economic, public health impacts and risk management options were made. Background information and data used for the risk analysis were summarised in the form of a separate knowledge document (Section 2.2).

2.2 Knowledge document

A literature search and data analysis describing the current body of knowledge with regard to taxonomy, habitat preference, dispersal mechanisms, current distribution, ecological and socio-economic impacts and management options for *L. major* were undertaken. The results of the literature search were presented in the form of a knowledge document (Matthews et al., 2012; Appendix 1) and distributed to an expert team in preparation for the risk assessment.

2.3 Risk assessment

2.3.1 Dispersion potential, invasiveness and ecological impacts

The ISEIA protocol assesses risks associated with dispersion potential, invasiveness and ecological impacts only (Branquart, 2007; ISEIA, 2009). The *L. major* risk assessment was carried out by an expert team. This team consisted of five individuals. One from the Netherlands Food and Consumer Product Safety Authority; one from the Dutch plant research and conservation organisation FLORON; one from the Roelf Pot Research and Consultancy firm and two from the Radboud University, Nijmegen. Each expert completed an assessment form independently, based on the contents of the knowledge documents. Following this preliminary individual assessment, the entire project team met, elucidated differences in risk scores, discussed diversity of risk scores and interpretations of key information. The results of these discussions were presented in an earlier draft of this report. Following the submission of this draft version to the expert team, further discussion led to agreement on consensus scores and the level of risks relating to the four sections contained within the ISEIA protocol (Table 2.1).
Table 2.1: Definitions of criteria for risk classifications per section used in the ecological risk assessment protocol (Branquart, 2007; ISEIA, 2009).

<table>
<thead>
<tr>
<th>1. Dispersion potential or invasiveness risk</th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Low</strong></td>
<td>The species does not spread in the environment because of poor dispersal capacities and a low reproduction potential.</td>
</tr>
<tr>
<td><strong>Medium</strong></td>
<td>Except when assisted by man, the species doesn’t colonise remote places. Natural dispersal rarely exceeds more than 1 km per year. However, the species can become locally invasive because of a strong reproduction potential.</td>
</tr>
<tr>
<td><strong>High</strong></td>
<td>The species is highly fecund, can easily disperse through active or passive means over distances &gt; 1km / year and initiate new populations. Are to be considered here plant species that take advantage of anemochory, hydrochory and zoochory, insects like Harmonia axyridis or Cemeraria ohridella and all bird species.</td>
</tr>
</tbody>
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<table>
<thead>
<tr>
<th>2. Colonisation of high conservation habitats risk</th>
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</tr>
</thead>
<tbody>
<tr>
<td><strong>Low</strong></td>
<td>Population of the non-native species are restricted to man-made habitats (low conservation value).</td>
</tr>
<tr>
<td><strong>Medium</strong></td>
<td>Populations of the non-native species are usually confined to habitats with a low or a medium conservation value and may occasionally colonise high conservation habitats.</td>
</tr>
<tr>
<td><strong>High</strong></td>
<td>The non-native species often colonises high conservation value habitats (i.e. most of the sites of a given habitat are likely to be readily colonised by the species when source populations are present in the vicinity) and makes therefore a potential threat for red-listed species.</td>
</tr>
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<table>
<thead>
<tr>
<th>3. Adverse impacts on native species risk</th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Low</strong></td>
<td>Data from invasion histories suggest that the negative impact on native populations is negligible.</td>
</tr>
<tr>
<td><strong>Medium</strong></td>
<td>The non-native is known to cause local changes (&lt;80%) in population abundance, growth or distribution of one or several native species, especially amongst common and ruderal species. The effect is usually considered as reversible.</td>
</tr>
<tr>
<td><strong>High</strong></td>
<td>The development of the non-native species often causes local severe (&gt;80%) population declines and the reduction of local species richness. At a regional scale, it can be considered as a factor for precipitating (rare) species decline. Those non-native species form long standing populations and their impacts on native biodiversity are considered as hardly reversible. Examples: strong interspecific competition in plant communities mediated by allelopathic chemicals, intra-guild predation leading to local extinction of native species, transmission of new lethal diseases to native species.</td>
</tr>
</tbody>
</table>

<table>
<thead>
<tr>
<th>4. Alteration of ecosystem functions risk</th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Low</strong></td>
<td>The impact on ecosystem processes and structures is considered negligible.</td>
</tr>
<tr>
<td><strong>Medium</strong></td>
<td>The impact on ecosystem processes and structures is moderate and considered as easily reversible.</td>
</tr>
<tr>
<td><strong>High</strong></td>
<td>The impact on ecosystem processes and structures is strong and difficult to reverse. Examples: alterations of physico-chemical properties of water, facilitation of river bank erosion, prevention of natural regeneration of trees, destruction of river banks, reed beds and / or fish nursery areas and food web disruption.</td>
</tr>
</tbody>
</table>
The ISEIA protocol contains twelve criteria that match the last steps of the invasion process (i.e., the potential for spread establishment, adverse impacts on native species and ecosystems). These criteria are divided over the following four risk sections: (1) dispersion potential or invasiveness, (2) colonisation of high conservation habitats, (3) adverse impacts on native species, and (4) alteration of ecosystem functions. Section 3 contains sub-sections referring to (i) predation / herbivory, (ii) interference and exploitation competition, (iii) transmission of diseases to native species (parasites, pest organisms or pathogens) and (iv) genetic effects such as hybridisation and introgression with native species. Section 4 contains sub-sections referring to (i) modifications in nutrient cycling or resource pools, (ii) physical modifications to habitats (changes to hydrological regimes, increase in water turbidity, light interception, alteration of river banks, destruction of fish nursery areas, etc.), (iii) modifications to natural successions and (iv) disruption to food-webs, i.e. a modification to lower trophic levels through herbivory or predation (top-down regulation) leading to ecosystem imbalance.

Each criterion of the ISEIA protocol was scored. Scores range from 1 (low risk) to 2 (medium risk) and 3 (high risk). Definitions for low, medium and high risk, according to the four sections of the ISEIA protocol are given in table 2.1. If knowledge obtained from the literature review was insufficient, then the assessment was based on expert judgement and field observation leading to a score of 1 (unlikely) or 2 (likely). If no answer could be given to a particular question (no information) then no score was given (DD - deficient data). Finally, the highest score within each section was used to calculate the total score for the species.

Consensus on the risk score of each section was reached using a hierarchical method where evidence from within the Netherlands was given priority over evidence derived from impacts occurring outside the Netherlands. It was also considered that the suitability of habitats in the Netherlands may change due to e.g. water temperature rise due to climate change. Moreover, consideration was given to the future application or non-application of management measures that will affect the invasiveness and impacts of this invasive plant in the Netherlands.

Subsequently, the Belgian Forum Invasive Species (BFIS) list system for preventive and management actions was used to categorise the species of concern (Branquart, 2007; ISEIA, 2009). This list system was designed as a two dimensional ordination (Environmental impact * Invasion stage; Figure 2.1). It is based on guidelines proposed by the Convention on Biological Diversity (CBD decision VI/7) and the European Union strategy on invasive non-native species. Environmental impact of the species was classified based on the total risk score (global environmental risk) which is converted to a letter / list: score 4-8 (C), 9-10 (B - watch list) and 11-12 (A - black list). This letter is then combined with a number representing invasion stage: (0) absent, (1) isolated populations, (2) restricted range, and (3) widespread.
Figure 2.1: BFIS list system to identify species of most concern for preventive and mitigation action (Branquart, 2007; ISEIA, 2009).

2.3.2 Socio-economic and public health impacts

Potential socio-economic and public health impacts did not form a part in the risk analysis according to the ISEIA protocol. However, these potential risks should be considered in an integrated risk analysis. Socio-economic risks were examined as part of the literature study (Matthews et al., 2012) and in discussions with project partners. Socio-economic risks occurring at present or in the future dependent on alterations in habitat suitability and management interventions were considered.

2.4 Risk management options

Management options were examined as part of the literature study and extensively described in the knowledge document (Appendix 1) and in discussions with project partners. A description of effective management options is given. These are specifically relevant to, and therefore recommended for, the Netherlands.
3. Risk analysis

3.1 Probability of arrival

Curly Waterweed (*Lagarosiphon major*) has been recorded in the Netherlands since 2003. However, the current distribution of *L. major* is characterised by isolated populations. Further introductions may result in a widened distribution of *L. major* in the Netherlands.

Brunel (2009) undertook a survey examining the importation of non-native aquatic plants to 10 countries in Europe. In 2006, the Netherlands imported some 20,000 units of *L. major*. These were used in aquaculture and garden ponds. The next most prolific importer of *L. major* was Germany where there were 5,200 records of import, however, this data was obtained over only 10 months in 2007 (Brunel, 2009). The increase in e-commerce has exacerbated the problem of invasive plant sale giving international retailers the ability to advertise online and send plants in the post (Kay & Hoyle, 2001).

A search of Google.nl, while not representative of the total current availability on the Dutch horticulture market, revealed a number of examples where *L. major* was advertised for sale on plant retailers websites (Figure 3.1). The results showed that the search term used had a large influence on the results found. Results obtained from *L. major* and Verspreidbladige waterpest contained no commercial websites and were mainly educational. However, results obtained using the term Gekroesde waterpest were biased towards retail and hobbyist websites. This indicates that measures applied to retailers and hobbyists must involve the use of all common names for *L. major* to avoid plants being sold under a name not used to educate retailers and the public. Also, the monitoring of retailers must involve the use of all commons names to avoid plants being missed.

![Figure 3.1](image_url)

*Figure 3.1:* Type of websites featuring Curly waterweed (*Lagarosiphon major*) found via Google.nl using different search terms.
Once bought, unwanted plants may be disposed of in the freshwater system. The results of a recent survey examining the behaviour of consumers of aquatic plants in the Netherlands showed that 2% of the 230 respondents had disposed of aquatic plants in open water (Verbrugge et al., 2011). Moreover, in Dutch waters, common garden pond plants occur with examples of pumpkinseed sunfish (*Lepomis gibbosus*). This fish species was introduced to the Netherlands in 1902 as an aquarium and garden pond fish (Van Kleef et al., 2008). This gives further credence to species disposal as a potential route for the introduction of invasive plants and animals. We predict that without management intervention, *L. major* introductions will continue, leading to potential increases in its distribution within the Netherlands. After considering the above information the probability of arrival was judged to be high.

### 3.2 Probability of establishment

*L. major* was first recorded in ditches in Soest when it occupied 4 kilometre squares (Valkenburg & Pot, 2008). However, the plants were located at a single location around the junction of four kilometre squares. The rate of dispersal peaked in the years 2007 and 2008. Since 2008 the number of reported records has decreased compared with preceding years. This, however, may well be an artefact as people no longer report the species for a particular site once it has been reported in preceding years. Moreover, a particular kilometre square may only be surveyed once every 5 years.

The current recorded distribution of *L. major* in the Netherlands is mainly characterised by isolated populations of plants (Figure 3.2). Some locations are, however, remotely interconnected by rivers and canals, such as in Drenthe and Groningen. In some cases the plant may have been overlooked in between the known stands. *L. major* grows in stagnant or slow-flowing water at water depths between 60-140 cm. At some sites the water is very turbid (Secchi disk readings less than 25 cm). All sites are situated in urban areas, although in Drenthe some sites are situated in rural areas (in Dutch called ‘veenwijken’) located close to urban areas.

In the ditches in Soest the plant population persisted for 9 years (A. Aptroot, personal communication). The plants survived last two, relatively severe winters in shallow water (30-70 cm) with almost no signs of damage. During the summer of 2012 the plants showed a strong growth. In some other areas, *L. major* has been present for a period of at least 5 years. In June 2012, *L. major* was observed at Emmer- Erfscheidenevenen, in the Musselkanaal and at Ter Apel. In Ter Apel the plants were only found in 2008 in high density. The same year the plants were removed partly in late summer to maintain the drainage function of the water body. Since 2009 the density of *L. major* was low and several other plant species were found. Some of them were locally abundant but none became dominant (personal communication J. Meeuse; Field observations and data Waterboard Hunze en Aa’s).
Figure 3.2: Distribution of Curly Waterweed (*Lagarosiphon major*) in the Netherlands since first introduction in 2003 (Data National Database Flora en Fauna, complemented with data sources mentioned Matthews *et al.*, 2012).

*L. major* displays a wide tolerance to different habitats and, in general, it is likely that conditions found in many lakes and low velocity streams and rivers found in the Netherlands will not prevent its establishment. The minimum temperature limit for survival of *L. major* is 10 °C. In the Netherlands surface water layer temperatures frequently fall below this level for an extended period in winter. Moreover, ice can form which has consequences for *L. major* living in turbid water as much of the biomass lies in the surface layer where light is accessible. Plants have to re-grow portions that are lost due to ice damage (Van Valkenburg, unpublished results). However, deeper warmer refuges may exist and the plant has been shown to sink in winter to avoid colder surface water (Centre for Ecology and Hydrology, 2004). Experiences from the UK show that in northern, colder areas, the plant mass collapses, but never dies down completely (Centre for Ecology and Hydrology 2004).

In New Zealand, *L. major* was found to exist in several habitats, spanning wide trophic, altitudinal and temperature ranges (de Winton *et al.*, 2009). *L. major* achieves its maximum vegetative expression in clear, still water and is tolerant of low nutrient conditions, but grows best in hard water with a good nutrient supply (Caffrey & Acavedo, 2007). In New Zealand, the rate of growth of *L. major* does not necessarily correlate with the trophic status or water chemistry of the waterbody (Brown & Dromgoole, 1977). However, in lakes with accelerated eutrophication and severely decreased water clarity,
*L. major* abundance declines (Coffey & Clayton, 1988). Inorganic carbon (as free CO$_2$), inorganic nitrogen and phosphorous are most important in controlling plant size (Riis et al., 2010). In conjunction with pH, *L. major* can survive in high alkalinity conditions as well (Invasive species compendium). *L. major* is sensitive to wave action and wind, preferring sheltered sites or reed beds that trap floating plant fragments (Caffrey & Acavedo, 2007). In Irish Lough Corrib, the plant was particularly abundant in sheltered, shallow bays and littoral areas. The plant was absent from rocky or boulder strewn locations within the lake and especially abundant where deep deposits of fine silt and organic mud accumulate (Caffrey & Acavedo, 2007) and is able to grow on sandy substrate where organic content is low and more coarse grained substrates (Chapman et al., 1971; Clayton et al., 1981; Caffrey & Acavedo, 2007). The probability of establishment in the Netherlands was judged to be low, due to 1) the limited recorded distribution of *L. major*, four years after being first recorded in the Netherlands, and 2) the relatively high sensitivity to low water temperature.

Table 3.1 gives an overview of the physiological tolerances of *L. major* identified during the literature search.

**Table 3.1:** Physiological conditions tolerated by Curly waterweed (*Lagarosiphon major*).

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Data origin</th>
<th>Physiological tolerance</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>Depth (m)</td>
<td>International</td>
<td>0.12 - 6.6$^6$</td>
<td>Coffey &amp; Wah (1988); Global Invasive Species Database (2007); Centre for Ecology and Hydrology (2004); Schutz (2008); Caffrey &amp; Acavedo (2007); Chapman et al. (1971)</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>International</td>
<td>10-25 (18-23 optimal)</td>
<td>Dutartre (1986); Australia Natural Heritage Trust (2003); GB Non-Native Species Secretariat (2011)</td>
</tr>
<tr>
<td>Temperature frost damage (°C)</td>
<td>International</td>
<td>-1$^a$</td>
<td>Bannister (1990)</td>
</tr>
<tr>
<td>Alkalinity (10$^{-3}$ eq/l)</td>
<td>The Netherlands</td>
<td>1.15-1.74$^2$</td>
<td>This study</td>
</tr>
<tr>
<td>pH</td>
<td>The Netherlands</td>
<td>6.5-7.0$^a$</td>
<td>This study</td>
</tr>
<tr>
<td>Light intensity (micro einsteins/m$^2$/h)</td>
<td>International</td>
<td>600 optimal</td>
<td>Schwarz &amp; Howard-Williams (1993)</td>
</tr>
<tr>
<td>Nitrate (mg/l)</td>
<td>International</td>
<td>1.05</td>
<td>Schutz (2008)</td>
</tr>
<tr>
<td>Phosphate (mg/l)</td>
<td>International</td>
<td>0.33</td>
<td>Schutz (2008)</td>
</tr>
</tbody>
</table>

$^a$: Lowest air temperature where no damage occurred (leaves exposed to air); $^b$: Non-light limited environments; $^c$: maximum for bicarbonate uptake; $^d$: See appendix 1 for results obtained from fieldwork.
3.3 Probability of spread

Outside its South African native range, only female plants are known (Cook, 1982; National Botanic Gardens, 2007). Reproduction and dispersal are facilitated by the detachment of small fragments and lateral branches that subsequently become rooted (Centre for Ecology and Hydrology, 2004). As all reproduction of *L. major* occurs through fragmentation or vegetatively, potential vectors that transfer plant fragments from colonised to uncolonised water bodies are of great importance (Table 3.2). Compton *et al.* (2012) linked the distribution of *L. major* in New Zealand lakes with human transport vectors. In this study high risk lakes lay in the vicinity of high human population densities, where lake access was relatively easy. Clayton *et al.* (1981) concluded that *L. major* was accidently introduced via boating vectors and observed that *L. major* distribution was associated with the most occupied, developed, and recreationally used area of Lake Rotoma, New Zealand. Establishment of vegetative fragments was often associated with fallen, submerged trees that had probably entangled drifting shoots (Clayton *et al.*, 1981). After considering the above information the probability of spread within the Netherlands was judged to be high.

Table 3.2: Potential dispersal vectors of Curly waterweed (*Lagarosiphon major*).

<table>
<thead>
<tr>
<th>Vector / mechanism</th>
<th>Mode of transport</th>
<th>Examples and relevant information</th>
<th>Importance to dispersal into and within the Netherlands</th>
<th>References</th>
</tr>
</thead>
<tbody>
<tr>
<td>Trade</td>
<td>Overland</td>
<td>E-commerce, plants transported in the post</td>
<td>High</td>
<td>Bowmer <em>et al.</em> (1995); Brunel (2009); GB Non-Native Species Secretariat (2011)</td>
</tr>
<tr>
<td>Hobbyists</td>
<td>Overland</td>
<td>Disposal of unwanted plants</td>
<td>High</td>
<td>Bowmer <em>et al.</em> (1995)</td>
</tr>
<tr>
<td>Boats / trailers</td>
<td>Upstream / downstream, overland</td>
<td>Occurs as a result of improper disinfection and moved from water body to water body</td>
<td>High</td>
<td>Bowmer <em>et al.</em> (1995); McGregor &amp; Gourley (2002)</td>
</tr>
<tr>
<td>Water current</td>
<td>Downstream</td>
<td>Plant fragments transported in flowing water</td>
<td>High</td>
<td>Bowmer <em>et al.</em> (1995)</td>
</tr>
<tr>
<td>Weed harvesters</td>
<td>Upstream / downstream, overland</td>
<td>Machinery not properly disinfected moved from water body to water body</td>
<td>Medium - high</td>
<td>McGregor &amp; Gourley (2002)</td>
</tr>
<tr>
<td>Fishing equipment</td>
<td>Upstream / downstream, overland</td>
<td>Occurs as a result of improper disinfection and moved from water body to water body</td>
<td>Medium</td>
<td>McGregor &amp; Gourley (2002)</td>
</tr>
<tr>
<td>Vehicles</td>
<td>Upstream / downstream, overland</td>
<td>Plants become trapped when crossing fords and subsequently transported</td>
<td>Low</td>
<td>Bowmer <em>et al.</em> (1995)</td>
</tr>
<tr>
<td>Large aquatic birds</td>
<td>Upstream / downstream, overland</td>
<td>Rare occurrence</td>
<td>Low</td>
<td>McGregor &amp; Gourley (2002); GB Non-Native Species Secretariat (2011)</td>
</tr>
</tbody>
</table>
3.4 Risk classification using the ISEIA protocol

3.4.1 Expert consensus scores

The risk classifications attributed to *L. major* for each section of the ISEIA protocol were medium or high (Table 3.3). The total risk score attributed to this species was 9 out of a maximum risk score of 12. This results in an overall classification of medium risk for this species.

Table 3.3: Consensus scores and risk classifications for Curly Waterweed (*Lagarosiphon major*) in the current situation in the Netherlands.

<table>
<thead>
<tr>
<th>ISEIA Sections</th>
<th>Risk classification</th>
<th>Consensus score</th>
</tr>
</thead>
<tbody>
<tr>
<td>Dispersion potential or invasiveness</td>
<td>high risk</td>
<td>3</td>
</tr>
<tr>
<td>Colonization of high value conservation habitats</td>
<td>medium risk</td>
<td>2</td>
</tr>
<tr>
<td>Adverse impacts on native species</td>
<td>medium risk</td>
<td>2</td>
</tr>
<tr>
<td>Alteration of ecosystem functions</td>
<td>medium risk</td>
<td>2</td>
</tr>
<tr>
<td>Global environmental risk</td>
<td>B - list category</td>
<td>9</td>
</tr>
</tbody>
</table>

3.4.2 Dispersion potential or invasiveness

Classification: **High risk.** *L. major* exhibits a strong reproduction potential and is able to reproduce vegetatively through fragmentation. Dispersal over a distance of greater than 1 kilometre per year may occur due to fragmentation and hydrochory. Moreover, *L. major* disperses via a variety of, mainly human, vectors. The importance of human vectors in the dispersal of this plant outside its native range is demonstrated by scattered distribution patterns that correlate with concentrations of human activity e.g. access points for vehicles and boats on lake shores. Moreover, there continues to be a strong market for *L. major* in the Netherlands demonstrated by the high number of plant imports and the availability of plants for sale online. The dispersal of *L. major* away from initial points of introduction in the Netherlands has been limited. Since 2003, 31 kilometre squares that contained *L. major* have been recorded. The majority of records are distant and isolated from each other, indicating multiple introductions. *L. major* is found in and around urban areas where it is likely that human introductions through the disposal of plants to the inland network of water bodies occur. Reasons for the observed limited dispersal after initial introduction maybe low water velocity or a lack of dispersal vectors at the colonised locations. There are a few examples where locations are remotely interconnected by rivers and canals, such as in the provinces of Drenthe and Groningen, where vegetative dispersal or dispersal facilitated by vectors may have occurred.
3.4.3 Colonisation of high conservation value habitats

Classification: Medium risk. *L. major* is found in and around urban areas and to date, no records exist in high conservation habitats in the Netherlands. However, a potential EU Habitats Directive type in which *L. major* may appear is H3150 (Natural eutrophic lakes with Magnopotamion or Hydrocharition-type vegetation). This habitat type features species like *Stratiotes aloides*, *Utricularia vulgaris* and *Hydrocharis morsus-ranae*. *L. major* can occur together with these species. A population of *L. major* is present in Soest, close to areas of peat-land containing plant species that are representative of a H3150 habitat type and the protected Soesterveen nature area. These areas are hydrologically connected to the water body containing *L. major* but have not been colonised. Regular management of the waterbody containing *L. major* occurs, which may encourage its dispersal. Therefore, it cannot be ruled out that the peat-land and protected areas around Soest may be colonised by *L. major* in the future.

3.4.4 Adverse impacts on native species

Classification: Medium risk. Impact criteria relating to predation and herbivory are not relevant for *L. major*. The result of the literature search revealed no information relating to the transmission of parasites and diseases. Moreover, impact criteria related genetic effects are not relevant for the Netherlands. Hybridisation or introgression with natives will not occur because closely related species are absent. Therefore, the risk classification is based on the competition sub-section. The major adverse impacts of *L. major* are related to interference and exploitation competition. Evidence relating to these impacts comes from foreign sources, however, it was considered that similar impacts could occur in the Netherlands if *L. major* was to become more widespread here. In Lough Corrib, Ireland and Lake Taupo, New Zealand a number of native species were lost following *L. major* invasion (Caffrey & Acavedo, 2007; Howard-Williams & Davies, 1988). Moreover, large weed beds of *L. major* attracted herbivorous birds and detritivores such as swans and crayfish which also adversely affect the native flora (Howard-Williams & Davies, 1988). In other locations, however, *L. major* has proven to be less aggressive. In some areas of New Zealand *L. major* has been displaced by other species and may co-exist with native species (McGregor & Gourlay, 2002). Heavy infestations confer no oxygen benefit on fish and other animals (Ramey, 2001). Herbivorous fish species may find *L. major* less palatable then native species. *L. major* was found to be least palatable to the Grass carp (*Ctenopharyngodon idella*) when compared to a group of 9 other species of waterweed (Edwards, 1974). Changes in habitat structure resulting from the dense canopy produced by *L. major* may impact indigenous fish species. Salmonids have a preference for open water conditions while the cyprinids, perch and pike commonly seek the cover provided by dense weed beds (Caffrey & Acavedo, 2007). Finally, significant changes in abundance and species composition within the macroinvertebrate community have been observed following invasion by *L. major* (Kelly & Hawes, 2005; Caffrey & Acavedo, 2007). However, other researchers have found no difference in the preference of macroinvertebrate groups between native macrophytes and *L. major* (Biggs & Malthus, 1982). To date, these type of effects have not been observed in the Netherlands. Field experimentation has
demonstrated that *L. major* is less competitive when grown under conditions found in the Netherlands than evidence from other countries suggests.

The minimum temperature limit for survival of *L. major* is 10 °C. In the Netherlands surface water layer temperatures frequently fall below this level for an extended period in winter. However, deeper warmer refuges may exist and the plant has been shown to sink in winter to avoid colder surface water (Centre for Ecology and Hydrology, 2004). Competitiveness may be reduced in very shallow water (maximum depth 30 cm) where freezing of the entire water compartment and top layer of the sediment occurs. Here, plants may suffer damage as they cannot sink away from the surface. The time of year at which plant growth is triggered will differ per location dependent on temperature. In other countries more southerly than the Netherlands, where the climate is warmer, growth occurs all year round (R. Pot, unpublished results). Limited growth due to low temperature will also impact competitiveness.

Reduced competitiveness and the lack of observations of impacts suggests a current low risk to native flora and fauna in the Netherlands. However, taking into account the high negative impact of *L. major* on native species seen in other countries in temperate regions, potential risks cannot be excluded in the Netherlands, resulting in a medium risk classification in this category.

Future increases in temperatures due to climate change and omitting measures to prevent human introduction may lead to an increase in the distribution and competitiveness of *L. major* in the Netherlands, also resulting in a greater impact on native species and an increase in the risk score.

### 3.4.5 Alteration of ecosystem functions

**Classification: Medium risk.** The risk classification is based on all four sub-sections contained within this section. Evidence of altered ecosystem functioning observed within the Netherlands is limited. Most of the evidence presented here is from foreign studies, however, it was considered that similar impacts could occur in the Netherlands if *L. major* was to become more widespread. A major impact of *L. major* on ecosystem functioning is light interception. Where mature surface-reaching stands have become established, the canopy is able to shade out, and competitively exclude, even tall submerged species (Figure 3.3). Changes in nutrient cycling, resource use and habitat structure in the presence of *L. major* result in increased dissolved reactive phosphorous and dissolved inorganic nitrogen and result in changes in temperature and dissolved oxygen level (Schwarz & Howard-Williams, 1993; Department of Primary Industries, 2011). Food webs involving fish species may be effected directly due to the change of species food source availability following *L. major* invasion (Edwards, 1974) and changes in the macroinvertebrate community (Kelly & Hawes, 2005; Caffrey & Acavedo, 2007). If the growth form of *L. major* is different from that of the aquatic plants that it replaces then changes in natural succession may occur.

However, many of the negative effects listed in the literature may be viewed as positive effects in the Netherlands. For example, increase in plant biomass and changes in nutrient cycling usually lead to a higher water transparency and more complex
invertebrate food web systems (Jeppesen et al., 1998). This is regarded as an improvement in water quality according to the water quality assessment of the EC Water Framework Directive (Van der Molen & Pot, 2007).

The lack of observed impacts to ecosystem functioning in the Netherlands suggests a current low risk in this category. However, the observed high impact on ecosystem functioning in other countries in temperate regions resulted in the experts involved coming to a consensus that the potential risk of the species is medium.

![Dense vegetation of Curly Waterweed (Lagarosiphon major) in a ditch near Ter Apel, the Netherlands (Photo: R. Pot).](image)

**Figure 3.3:** Dense vegetation of Curly Waterweed (*Lagarosiphon major*) in a ditch near Ter Apel, the Netherlands (Photo: R. Pot).

Future increasing water temperatures due to climate change may have an impact on ecosystem functions e.g. modification of natural resources and disruption to food-webs that may lead to an increase in the risk score in the future.

### 3.4.6 Species classification

The species classification corresponds to the global environmental risk score of the ISEIA (Table 3.1) combined with the current distribution of the non-native species within the country in question. The species classification for *L. major* is B1 (Figure 3.4). This indicates a non-native species with isolated populations and moderate environmental hazard (i.e. ecological risk) that should be placed on a watch list.

![Curly Waterweed (Lagarosiphon major) classification according to the BFIS list system.](image)

**Figure 3.4:** Curly Waterweed (*Lagarosiphon major*) classification according to the BFIS list system.
However, habitat alteration resulting from climate change may result in a future re-grading of risk. Future increases in the area suitable for the colonisation of *L. major* due to increased water temperature during winter periods may occur. According to climate change scenarios developed by the Dutch Royal Meteorological Institute (KNMI; Van den Hurk et al., 2006) the average air temperature during winter may increase by 0.9-2.3 °C and 1.8-4.6 °C over the years 1990-2050 and 1990-2100. This means that water temperatures during winter periods will also increase over these periods. Therefore, it is probable that *L. major* growth and survival will be limited to a lesser degree by minimum water temperature in future. This particularly holds for deeper growing sites in clear water bodies. Moreover, if male plants are imported in conjunction with the current importation of female plants, then *L. major* may increase its dispersal and colonisation potential through the formation of seeds and associated increase in genetic vigour. In the absence of management intervention, these developments may result in wider distributions of *L. major* dependent on the availability of dispersal vectors. Colonisation of high conservation value habitats could potentially occur if they were accessed by vectors such as pleasure boats or anglers. Wider distributions of *L. major* will likely result in high impacts on native species and wide alterations to ecosystem functions. This would lead to a reclassification to high risk according to the ISEIA protocol (Table 3.4). In this theoretical scenario the distribution of *L. major* would increase to at least a restricted range resulting in at least an A2 classification.

**Table 3.4:** Curly Waterweed (*Lagarosiphon major*) theoretical classification according to potential future habitat scenario.

<table>
<thead>
<tr>
<th>ISEIA Sections</th>
<th>Risk classification</th>
<th>Consensus score</th>
</tr>
</thead>
<tbody>
<tr>
<td>Dispersion potential or invasiveness</td>
<td>high risk</td>
<td>3</td>
</tr>
<tr>
<td>Colonization of high value conservation habitats</td>
<td>medium risk</td>
<td>2</td>
</tr>
<tr>
<td>Adverse impacts on native species</td>
<td>high risk</td>
<td>3</td>
</tr>
<tr>
<td>Alteration of ecosystem functions</td>
<td>high risk</td>
<td>3</td>
</tr>
<tr>
<td>Global environmental risk</td>
<td>A - list category</td>
<td>11</td>
</tr>
</tbody>
</table>

### 3.5 Socio-economic impacts

Due to limited distribution of *L. major* in the Netherlands, the current socio-economic impact of the species is expected to be low. However, provincial water-boards are having to manage local populations of *L. major* due to water-flow obstruction at some locations. Potential future changes as a result of e.g. a rise in water temperature due to climate change, may increase the suitability and area of *L. major* habitat. Socio-economic impacts that have occurred in other countries are considered as it is possible that these impacts may occur in the Netherlands in the future.

In its South African native range as well as in introduced areas, prolific growth of *L. major* can interfere with commercial navigation and water-based recreation (Centre for Ecology and Hydrology, 2004; Caffrey & Acavedo, 2007). Swimming maybe impossible in areas of dense weed growth and the snarling of weeds in outboard motors may put recreational boaters at risk (Caffrey & Acavedo, 2007). Storms can tear the weed loose and deposit large masses of rotting vegetation on beaches, spoiling their amenity value;
and effect power stations (Brown, 1975; Rowe and Hill, 1989). Large beds of \textit{L. major} may increase the risk of flow impedance as the discharge capacity of an invaded water body is reduced (Department of Primary Industries, 2011). Extensive growth can block the turbine screens of hydro-electric power stations in quantities too great for the cleaning machinery to clear, causing temporary shutdowns, economic losses and power shortages (Chapman \textit{et al.}, 1974).

In the United Kingdom, controlling \textit{L. major} by mechanical means was estimated to cost 1,250 Euros per hectare per year assuming that each 10 km square contains at least 1 hectare of plants (GB Non-Native Species Secretariat, 2011).

### 3.6 Public health effects

During the literature study or in communications with project partners there was no information found concerning the public health effects of \textit{L. major}.

### 3.7 Risk management options

#### 3.7.1 Prevention

The main distribution channel or vector is trade of plants for aquaria and garden pools. One possible solution is to trade in a native plant or a non-native one with a low potential impact that replaces \textit{L. major}. The best alternative native species is \textit{Ceratophyllum demersum}. Another alternative for \textit{Curly Waterweed} (\textit{Lagarosiphon major}) for trade may be \textit{Elodea nuttallii}. This is also a non-native species, but it has established and has already become very common in the Netherlands. Negative impacts relating to \textit{E. nuttallii} are potentially already present and the countrywide removal of this species is not a feasible option. Due to its already widespread distribution, it is not expected that an increase in the severity of impacts will occur following further introductions of \textit{E. nuttallii} within the Netherlands. Currently, in the Netherlands, a campaign is underway that aims to prevent further introductions and spread by making consumers and employees from garden centres and plant nurseries more aware of the problems with non-native species. The name of this campaign is ‘Geen exoot in de sloot’. Its effectiveness is currently being examined (Verbrugge \textit{et al.}, 2010).

Public awareness is an important component in a strategy aimed at controlling or removing an invasive species from a catchment area. This is especially important in the Netherlands due to the high level of imports and trade of \textit{L. major} associated with the fact that people are the major vector for dispersal of this species. Awareness leaflets, press releases, calendars, lakeside notifications and an information website, warning of the environmental, economic and social hazards posed by this plant will contribute to public awareness (Caffrey & O’Callaghan, 2007).

Education of anglers and boaters may be especially useful as they can assist in reporting sightings of the plant. Moreover, instruction on the decontamination of boating and angling equipment is necessary to prevent dispersal of \textit{L. major}. A guide for the identification of aquatic invasive species, describing associated impacts and strategies
for prevention of spread was produced in the Netherlands in conjunction with the ‘Code of conduct on aquatic plants’ (Van Valkenburg, 2011). Its aim is to create awareness and assist in the monitoring of non-native aquatic plants.

3.7.2 Elimination

Once populations of the plants have established, eradication is very difficult. The best option is to isolate the local populations and intervene as little as possible. At the very least a natural lowering of fitness and abundance may be expected, as was previously observed in Ter Apel, the Netherlands. Here, manual removal of plants to prevent blocking discharge flow occurred only in the first year of appearance in 2008 (Figure 3.1). The plants did not recover to the same density in following years.

3.7.3 Control

If active control of *L. major* is required, as in Emmer-Erfscheidenveen, the Netherlands, removal using weed cutting machinery is recommended e.g. mowing baskets or weed cutting boats, and the prevention of fragment spread (Figure 3.5).

![Figure 3.5: A weed cutting boat with adjustable mowing gear used for aquatic weed control in the Netherlands (Photo: R. Pot).](image)

However, mechanical methods may result in the breakup of plant stems resulting in the dispersal of plants to new areas (Bowmer *et al.*, 1995). The dispersal of plant fragments and subsequent vegetative reproduction has been observed following mechanical harvesting in the Netherlands (R. Pot, unpublished results). Plants spread to connected water bodies after cutting at Emmer-Erfscheidenveen. When management methods were not introduced, the plants did not spread at most of the known sites in the Netherlands. Therefore, the best method to prevent spread of the species is to show reticence when considering the implementation of management.
4. Discussion

4.1 Gaps in knowledge and uncertainties

A lack of information in the literature on the (potential) impact of Curly Waterweed (*Lagarosiphon major*) in the Netherlands has resulted in a reliance on expert knowledge and field observations to judge the level of certain impacts. This lack of information may be a reflection of the limited distribution of *L. major* in the Netherlands at the present time.

The high risk associated with *L. major* to native species and ecosystem functions in other countries maybe a function of a greater habitat suitability and resultant high level of invasiveness in those countries.

Future changes such as increases in water temperature associated with climate change may result in an increase in the distribution of *L. major* in the Dutch freshwater network as well as in isolated water bodies. Therefore, the risk of impacts may have to be reassessed in future in view of greater potential impacts.

The ISEIA protocol is limited to an assessment of invasiveness and ecological impacts. No assessment of socio-economic impacts or impacts to human health are considered and are not considered in the calculation of global environmental risk score. Socio-economic impacts or impacts to human health were therefore considered separately.

Risk criteria in the ISEIA protocol were sometimes restrictive, as there was an absence of quantitative data that allowed the criteria to be assessed e.g. 1 km per year dispersal criterion for the ‘dispersion or invasiveness’ section.

4.2 Comparison of available risk classifications

The ISIEA protocol has been used to assess the risk of *L. major* in Belgium. *L. major* was classified as high risk in all four ISEIA categories.

A number of other assessments have been used in other countries and have consistently scored *L. major* as a high risk invasive species. In the United Kingdom, application of the UK Risk Assessment Scheme resulted in *L. major* being given a high risk rating (GB Non-Native Species Secretariat, 2011). In Ireland *L. major* is defined as high risk, scoring 20 in the IS Ireland risk assessment (Irish National Invasive Database, 2007). In this assessment low risk species score 0-12, medium risk species 13-19 and high risk species greater than 19. In Spain, *L. major* scored 18 on a scale ranging from -14 to 30 on the Weed Risk Assessment protocol (WRA). Species scoring over 6 in the WRA are rejected for introduction due to their potential impacts (Andreu & Vilà, 2010). Finally, in Australia, the Victorian Weed Risk Assessment (WRA), while not giving an overall score, categorised *L. major* as high risk for adverse impacts to water quality and native plant species resulting from structural habitat change (Department of Primary Industries, 2011).
The medium risk score obtained from this assessment is therefore at odds with assessments carried out in other countries. High impacts have not been observed in the Netherlands, even though the species has been present since 2003. The main reasons for this are 1) the limited recorded distribution of L. major, 2) possible lack of vectors at locations where the species was introduced, and 3) minimum temperature tolerance (> 10 °C) in relation to current water temperatures (< 10°C) during winter periods (species-environment mismatch). However, monitoring of L. major should continue due to potential future increases in population, due to increasing water temperature during winter periods caused by climate change, which may result in higher levels of impact.

4.3 Risk management

Banning of sale of invasive plants via the plant trade and creation public awareness of consumers continue to be the most potentially effective methods of controlling introduction and the spread of invasive plant species. Based on current dispersion and potential invasiveness and risk it is recommended that L. major plants are sold with information that informs buyers of their potential invasive nature and the circumstances within which the plant can safely be used.

Once L. major is released to the environment, control and elimination becomes more difficult. Management by mechanical means has been recommended as management measure for control and possible elimination of the species. However, managers may first wish to consider observing the dispersal potential of individual populations of L. major prior to instigating active measures. If populations become problematic, isolation may be considered as this will facilitate the elimination of the species. The costs and risk of a facilitation of dispersal together with the ongoing limited distribution of L. major observed in the Netherlands since 2003, count against the early implementation of weed cutting as a control measure.
5. Conclusions and recommendations

The main conclusions and recommendations of the Risk analysis of non-native Curly Waterweed (*Lagarosiphon major*) in the Netherlands are as follows:

- The probability of *L. major* arriving in the Netherlands is determined largely by the plant trade and the swapping of plants between hobbyists. A number of internet websites were found that featured traders who were advertising *L. major* for sale within the Netherlands under the Dutch common name ‘Gekroesde waterpest’.

- We predict that without management intervention, *L. major* introductions will continue, leading to potential increases in its distribution within the Netherlands.

- The probability of arrival in the Netherlands was judged to be high.

- The current recorded distribution of *L. major* in the Netherlands is mainly characterised by isolated populations of plants.

- Due to the limited recorded distribution of *L. major*, nine years after being first recorded in the Netherlands, the probability of establishment within the Netherlands was judged to be low.

- Due to the lack of female plants in the Netherlands, reproduction and secondary spread are facilitated by the detachment of small fragments and lateral branches. These fragments subsequently become rooted, developing into new plants.

- In the Netherlands, secondary spread of fragments may be facilitated by dispersal vectors. Vectors can be ordered in terms of importance: hobbyists, boats and water flow (high); weed harvesters (medium-high), fishing equipment (medium); vehicles, large aquatic birds (low).

- The limited distribution of *L. major* in and around urban areas and evidence showing that a small proportion of hobbyists dispose of plants into the freshwater network, suggests that voluntary introductions by the public may be the major pathway through which *L. major* reaches the freshwater network in the Netherlands.

- The dispersal of *L. major* away from its initial points of introduction in the Netherlands has been limited and its distribution is characterised by isolated populations. This may be due a lack of (secondary) dispersal vectors e.g. water-flow, boats. Moreover, colonised sites are predominantly located in water bodies that are hydrologically isolated.

- The probability of spread within the Netherlands was judged to be high.

- *L. major* is found in and around urban areas and no records exist in high conservation value habitats in the Netherlands.
While there are many examples of impacts on native species observed in other countries, no effects have been observed in the Netherlands. Future changes to habitat resulting from e.g. climate change, may result in a revision of this risk score in the future.

There is limited evidence demonstrating negative impacts on the functioning of ecosystems in the Netherlands. However, in other countries featuring milder climates, where L. major has become more widely established, negative impacts on ecosystem functioning has been extensive. Future changes to habitat resulting from e.g. climate change, may result in a revision of this risk score in the future.

L. major was rated as a medium risk species for ecological impacts according to the ISEIA protocol. Its current limited distribution in the Netherlands, combined with this medium risk score, results in a B1 classification in the BFIS list system. Future changes in habitat characteristics due to climate change may result in greater (ecological) impacts as a result of increases in L. major distribution and a reclassification to A2 in the BFIS list system.

Due to limited distribution of L. major in the Netherlands, the current socio-economic impact of the species is expected to be minimal. However, potential future changes as a result of e.g. a rise in water temperature due to climate change, may increase the suitability and area of L. major habitat.

There was no information found concerning the public health effects of L. major during the literature study or in communications with project partners.

Managers may first wish to consider observing the dispersal potential of individual populations of L. major prior to instigating active measures. Costs and the risk of a facilitation of dispersal together with the limited dispersal of L. major observed in the Netherlands since 2003, count against the early implementation of weed cutting as a control measure.

If populations become problematic, isolation may be considered as this will facilitate the elimination of the species. If active management will be required (e.g. in case of obstruction of water discharge in drainage ditches), mechanical means have been recommended for control and possible elimination of L. major.

Based on current risk score B1 and expected increase in dispersion, potential invasiveness and risk in future due to increasing water temperatures during winter periods by climate change, it is recommended that L. major plants are sold with information that informs buyers of their potential invasive nature and the circumstances within which the plant can safely be used.
6. Acknowledgements

We thank the Netherlands Food and Consumer Product Safety Authority (Office for Risk Assessment and Research, Invasive Alien Species Team) of the Dutch Ministry of Economic Affairs, Agriculture and Innovation) for financial support of this study and all volunteers and organisations that contributed to the knowledge report on *L. major*. Dr. Trix Rietveld-Piepers and Wiebe Lammers of the Invasive Alien Species Team delivered constructive comments on an earlier draft of this report.
7. References


Appendices

Appendix 1. Knowledge document used for the risk analysis